

Research Article

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Co₃O₄-ZnO/rGO catalyst preparation and rhodamine B degradation by sulfate radical photocatalysis

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Abstract: The development of a combined photocatalytic system with peroxymonosulfate (PMS) has great potential applications in the degradation and treatment of aqueous organic pollutants. Herein, a Co₃O₄-ZnO/rGO was prepared by a hydrothermal method using cobalt acetate, zinc acetate, and reduced graphene oxide (rGO) as the main raw materials. The physical and chemical characteristics of the obtained catalyst were analyzed using X-ray diffraction (XRD), X-ray photoelectron spectroscopy (XPS), scanning electron microscopy (SEM), and Fourier transform infrared (FT-IR). The photocatalytic features and capacities of the catalytic materials to activate PMS were investigated. Co₃O₄-ZnO/rGO exhibited stronger photocatalytic activity and ability to activate PMS than Co₃O₄/rGO or ZnO/rGO, and significantly improved the ability of PMS and photocatalysis to synergistically degrade rhodamine B (RhB), with a degradation rate of 90.40% within 40 min. The mechanism of RhB degradation was proposed based on characterization of materials, evaluation of RhB degradation efficiency, and analysis of the active species involved. The unique particle/sheet structure of Co₃O₄-ZnO/rGO provides a large number of active sites, and the formation of heterojunctions between Co₃O₄ and ZnO improves carrier separation and transport in the reaction system. Our study offers a reference for designing more effective heterojunction catalysts based on the combination of PMS and photocatalytic technology.

Key words: Co₃O₄-ZnO/rGO catalyst; Rhodamine B (RhB); Heterojunction; Photocatalysis; Peroxymonosulfate (PMS)

1 Introduction

Over the past decade, the gradual intensification of water pollution has become a major concern (Yang ZH et al., 2019). Rhodamine B (RhB) is used extensively as a synthetic basic dye in numerous applications in the textile, leather, and plastic industries. However, RhB is chemically stable, tends to accumulate in the environment, is highly toxic and carcinogenic, and causes harm to humans and aquatic organisms even at low concentrations in wastewater. Advanced oxidation processes utilize strongly oxidizing reactive free radicals to effectively degrade various pollutants and are widely used in environmental treatments because of their high oxidation efficiency and fast degradation

rate (Ghanbari and Moradi, 2017; Cao et al., 2020). Compared with traditional technologies, photocatalytic technology has attracted increasing attention because of its environmental remediation potential and low cost (Gautam et al., 2020; Malefane et al., 2020). However, because they are limited by the narrow absorption range of ultraviolet (UV) light, the catalysts have low light-energy utilization efficiency in practical applications, which affects their photocatalytic degradation potential. Moreover, photocatalytic systems alone are not ideal for the treatment of polluted water bodies with low transmittance and high turbidity. In recent years, water pollution remediation has been greatly aided by the application of sulfate-radical-mediated photocatalysis (SR-photo) technology, which relies on the synergistic effects of photocatalysis and activated peroxymonosulfate (PMS) (Ren et al., 2018; Guo et al., 2020; Han et al., 2020). Under the influence of light and the catalyst in the SR-photo system, PMS can also produce other active species, such as sulfate (SO₄^{•-})

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and hydroxyl radicals ($\cdot\text{OH}$). Additionally, it can increase the effectiveness of photocatalytic oxidation reaction and decrease recombination of electron-hole pairs during the photocatalytic process (Zhang JL et al., 2020, 2021). However, this technique relies on the performance of the selected catalysts in terms of both their photocatalytic and PMS-activation properties. Therefore, finding a reliable and effective photocatalyst to activate PMS is crucial for the water-treatment industry.

Co_3O_4 is a p-type semiconductor with a narrow band gap (1.2–2.1 eV) and is widely used in photocatalysis because of its excellent stability and powerful light-absorption properties (Tang et al., 2016; Ren et al., 2020). In addition, studies have shown that Co (II) has a good activation effect on the PMS system (Yang J et al., 2019; Shen et al., 2020). However, as research on the photocatalytic efficacy of Co_3O_4 has expanded in recent years, photogenerated electrons of pure Co_3O_4 have been found to bind easily, which reduces its photocatalytic capacity (Guo et al., 2021). Moreover, the low Co_3O_4 redox potential (2.44 eV) renders the use of large quantities of Co_3O_4 as a photocatalyst difficult with regard to environmental purification (Tang et al., 2016). Numerous studies have shown that p-n heterojunctions formed by combining p- and n-type semiconductors can significantly reduce the compounding of photogenerated electron-hole pairs and enhance photocatalytic activity (Reda et al., 2017; Guo et al., 2018). ZnO is regarded as a suitable photocatalytic material and as an n-type semiconductor because of its nontoxicity, relatively low cost, and strong photochemical reactivity. At the p-n junction interface, the presence of p-type Co_3O_4 and n-type ZnO creates an advantageous internal electric field for photocatalysis, which may significantly decrease recombination of photogenerated electrons and holes and increase the efficiency of photoreactions (Dong et al., 2015).

Reduced graphene oxide (rGO) is an excellent semiconductor with a large pore size, specific surface area, and high electrical conductivity. It is widely used in the adsorption and oxidation of organic pollutants in wastewater and is widely studied as a carbon-based carrier (Li and Dong, 2019; Thakur and Kandasubramanian, 2019). Overall, the introduction of rGO into photocatalysts can significantly improve the photocatalytic activity of materials to enhance the efficient degradation of organic pollutants (Akbarzadeh et al., 2020). To our knowledge, no reports exist regarding the use of rGO-loaded Co_3O_4 -ZnO heterojunction as a dual-effect

catalyst to promote the degradation of RhB by a synergistic system of PMS and photocatalytic oxidation.

In this study, we investigated the capacity of a Co_3O_4 -ZnO/rGO heterojunction catalyst produced by a hydrothermal method to degrade RhB in solution by activating PMS under UV light. In addition, the effects of different preparation conditions, such as the rGO doping amount, hydrothermal reaction time, and calcination temperature, on the catalyst performance were investigated. The primary active compounds generated throughout the degradation process and the synergistic effect of the UV/ Co_3O_4 -ZnO/rGO/PMS system on the RhB removal mechanism were investigated. This provides a feasible strategy for constructing efficient and reusable dual-effect heterojunction catalysts.

2 Materials and methods

2.1 Chemicals

rGO was supplied by The Sixth Element (Changzhou) Materials Technology Co. Ltd. (Changzhou, China). Potassium PMS was obtained from the Chengdu Kelong Chemical Reagent Factory (Chengdu, China). Other analytical grade reagents including sulfuric acid (H_2SO_4), sodium nitrate (NaNO_3), potassium permanganate (KMnO_4), hydrochloric acid (HCl), cobalt acetate ($\text{Co}(\text{C}_2\text{H}_3\text{O}_2)_2$), zinc acetate ($\text{Zn}(\text{C}_2\text{H}_3\text{O}_2)_2$), anhydrous ethanol, ammonia ($\text{NH}_3\text{-H}_2\text{O}$), tert-butanol (TBA), triethanolamine (TEOA), methanol (MeOH), and benzoquinone (BQ) were purchased from Chongqing Chuandong Chemical Co., Ltd. (Chongqing, China). RhB was purchased from Tianjin Guangfu Chemical Co. Ltd. (Tianjin, China).

2.2 Synthesis

We used a hydrothermal approach to generate the Co_3O_4 -ZnO/rGO employed in the tests, and the precise processes are illustrated in Fig. 1.

First, 5-g rGO was dispersed in 15-mL deionized water, and the suspension was sonicated for 30 min. Cobalt acetate (0.2982 g) was weighed and added to 15-mL deionized water, stirred for 15 min to dissolve completely, and then mixed with the rGO suspension, after which 20-mL $\text{NH}_3\text{-H}_2\text{O}$ was added to the mixed solution and stirred for 30 min. The mixed solution was condensed and refluxed in an oil bath at 100 °C for 4 h, and the bottle was opened to volatilize

$\text{NH}_3\text{-H}_2\text{O}$ and then labeled as sample #1. Then, 0.505-g $\text{Zn}(\text{CH}_3\text{COO})_2$ was dissolved in 20-mL deionized water, 5-mL $\text{NH}_3\text{-H}_2\text{O}$ was added, and the mixture was condensed and refluxed for 2 h and labeled as sample #2. Afterward, samples #1 and #2 were mixed well, transferred into an autoclave, and reacted at 180 °C for 20 h. The suspension was then separated by centrifugation at 3000 r/min for 5 min, washed three times with deionized water and anhydrous ethanol, and dried at 60 °C for 8 h. Next, the powder was calcined at 500 °C for 2 h in a nitrogen-filled tube muffle furnace to obtain $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$ as the final product.

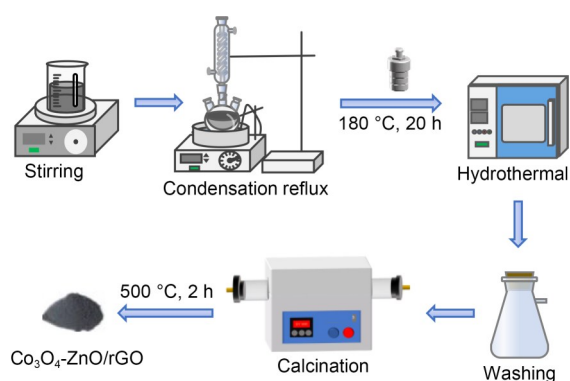


Fig. 1 Synthesis path of $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$

2.3 Characterization

X-ray photoelectron spectroscopy (XPS) studies of surface chemical morphology were performed using a K-Alpha photoelectron spectrometer (Thermo Fisher Scientific, USA) with monochromatic Al $K\alpha$ radiation at 15 kV. We obtained Fourier transform infrared (FT-IR) spectra using a Nicolet 670 spectrometer (Nicolet, USA) and analyzed the functional groups and chemical bonds of the materials using KBr as a reference in the spectral range of 4000–500 cm^{-1} . X-ray diffraction (XRD) curves were obtained using an X'Pert PRO instrument (PANalytical, the Netherlands) with Cu $K\alpha$ radiation with a 2θ range of 0°–90° and a sweep of 10°–90° at a rate of 5°/min. Sample morphology was examined by scanning electron microscopy (SEM) (ZEISS Gemini 300, Oberkochen, Germany), and the optical properties were determined by ultraviolet-visible diffuse reflectance spectrometry (UV-Vis DRS) (UV-3600, Shimadzu, Japan).

2.4 Experimental procedure

First, we examined the performance of the catalyst in relation to the rGO doping amount, duration of

the hydrothermal reaction, and calcination temperature. $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$ with different rGO doping fractions (0.1%, 0.3%, 0.5%, 0.7%, and 0.9%) was added to 50-mg/L RhB, after which the catalytic performance was evaluated in terms of the RhB degradation rate. Similarly, we tested the catalytic performance of $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$ prepared at hydrothermal times of 3.0, 3.5, 4.0, 4.5, and 5.0 h and calcination temperatures of 300, 400, 500, 600, and 700 °C.

The following three reaction systems were built: UV/ $\text{Co}_3\text{O}_4\text{-ZnO/rGO/PMS}$, UV/ $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$, and $\text{Co}_3\text{O}_4\text{-ZnO/rGO/PMS}$. Beakers #1, #2, and #3 each contained 200-mL RhB solution (50 mg/L). Then, 0.125-g $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$ was added, and the mixture was magnetically stirred for 45 min under dark conditions. After that, 0.025-g PMS was added to beakers #1 and #2; beakers #1 and #3 were irradiated with UV, while beaker #2 remained in the dark, and reactions were allowed to continue in the three beakers for 1 h. Samples were collected regularly. Immediately after sampling, 1-mL MeOH was added to stop the reaction. The samples were centrifuged at 3000 r/min for 5 min, the supernatant was quickly collected, and a UV spectrophotometer was used to detect RhB absorbance at 554 nm, which was then converted into RhB concentration.

Furthermore, we investigated the stability and reusability of $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$. After each experiment, the remaining catalyst was collected by centrifugation and washed three times with deionized water and ethanol and then dried at 60 °C for 8 h. While keeping the reaction conditions consistent, the experiment was repeated five times. The reusability of the material was determined based on its RhB degradation rate. The catalysts were characterized by XPS before and after the experiments, and the stability of the materials was determined by changes in the Co, Zn, and GO peak positions and corresponding intensities.

3 Results and discussion

3.1 Characterization

3.1.1 XRD

The XRD pattern in Fig. 2 reveals that the peak of $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$ had two components, Co_3O_4 (JCPDS, No. 43-1003) and ZnO (JCPDS, No. 36-1451) (Han et al., 2014; Lou et al., 2018). Peaks were observed at

19.3°, 34.4°, 44.7°, 54.9°, and 63.2° for 2θ , corresponding to the reflections of (111), (311), (400), (422), and (440), respectively. These corresponded well with the characteristic peaks of Co_3O_4 (Jin et al., 2020b). Similarly, the peaks corresponding to the characteristic peaks of ZnO were at 31.6°, 36.4°, 56.4°, 61.3°, and 67.7° for 2θ , corresponding to the (002), (102), (110), (103), and (220) reflections, respectively. Other characteristic peaks were also observed in the spectrum of Co_3O_4 -ZnO/rGO, indicating the presence of impurities in the synthesized product. The peaks of Co_3O_4 and ZnO in Co_3O_4 -ZnO/rGO produced a certain angle shift, which may have been caused by Co or Zn ions entering each other's lattices (Alshaikh et al., 2021).

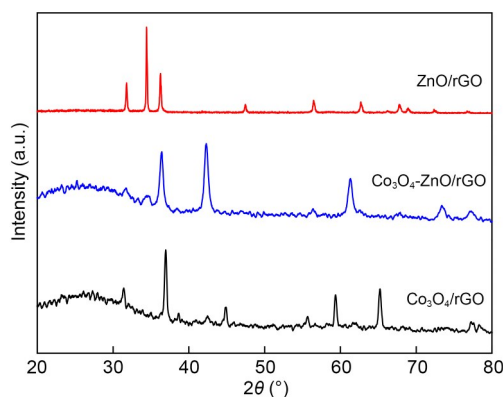


Fig. 2 XRD patterns of ZnO/rGO, Co_3O_4 -ZnO/rGO, and Co_3O_4 /rGO

3.1.2 FT-IR

Fig. 3 shows the FT-IR spectra of the three prepared materials. Two peaks can be observed at 572 and 667.5 cm^{-1} for Co_3O_4 /rGO, corresponding to the vibrations of Co (II)-O and Co (III)-O, respectively. In the spectrum of ZnO/rGO, the peak at 483 cm^{-1} corresponds to the stretching vibration of Zn (II)-O (El-Molla et al., 2017; Yang J et al., 2019). All the characteristic peaks of the above two components are observable in the Co_3O_4 -ZnO/rGO spectrum. In addition, the peak at 1634 cm^{-1} corresponds to the stretching vibration of the C=C sp^2 hybrid carbon atoms in reduced graphene oxide. The vibration of the water molecule-OH adsorbed on the catalyst surface was responsible for the absorption peak at 3419.5 cm^{-1} (Rakibuddin and Ananthakrishnan, 2016). Combined with the XRD characterization, the results showed that the material contained Co-O and Zn-O.

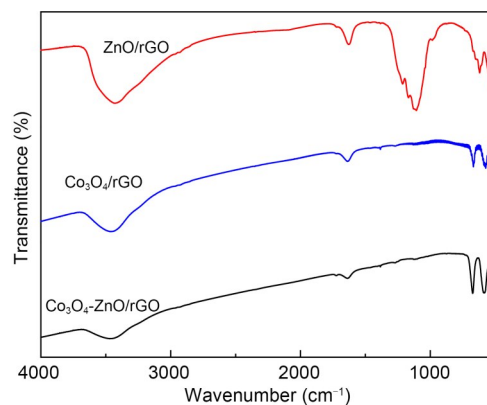


Fig. 3 FT-IR spectra of ZnO/rGO, Co_3O_4 /rGO, and Co_3O_4 -ZnO/rGO

3.1.3 SEM

The structure and morphology of the prepared material particles were characterized through sample SEM imaging. As shown in Fig. S1 of the electronic supplementary materials (ESM), Co_3O_4 exhibited a cubic crystal structure, aggregated and attached to a lamellar structure with folds. ZnO exhibited a hexagonal crystal structure attached to the sheet structure in the form of rods. The Co_3O_4 -ZnO/rGO image in Fig. 4 shows particles of different shapes with sizes ranging from 30 to 200 nm tightly attached to the lamellar structure; some are in contact and bonded together, indicating the formation of heterojunctions.

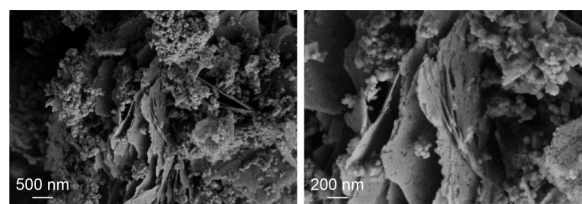


Fig. 4 SEM image of Co_3O_4 -ZnO/rGO

3.1.4 XPS

XPS was used to evaluate the elemental composition of the material surface. The results are shown in Fig. 5. For Co_3O_4 -ZnO/rGO, we observed peaks corresponding to C 1s, O 1s, Zn 2p, and Co 2p. In the C 1s XPS spectrum, Co_3O_4 -ZnO/rGO had two distinct peaks at 284.8 and 286.5 eV, which were assigned to the C=C and C-O lattice absorption peaks in graphene, respectively (Rakibuddin and Ananthakrishnan, 2016). Two peaks are shown at O 1s for Co_3O_4 -ZnO/rGO, with 530.4 and 529.9 eV corresponding to the Co_3O_4

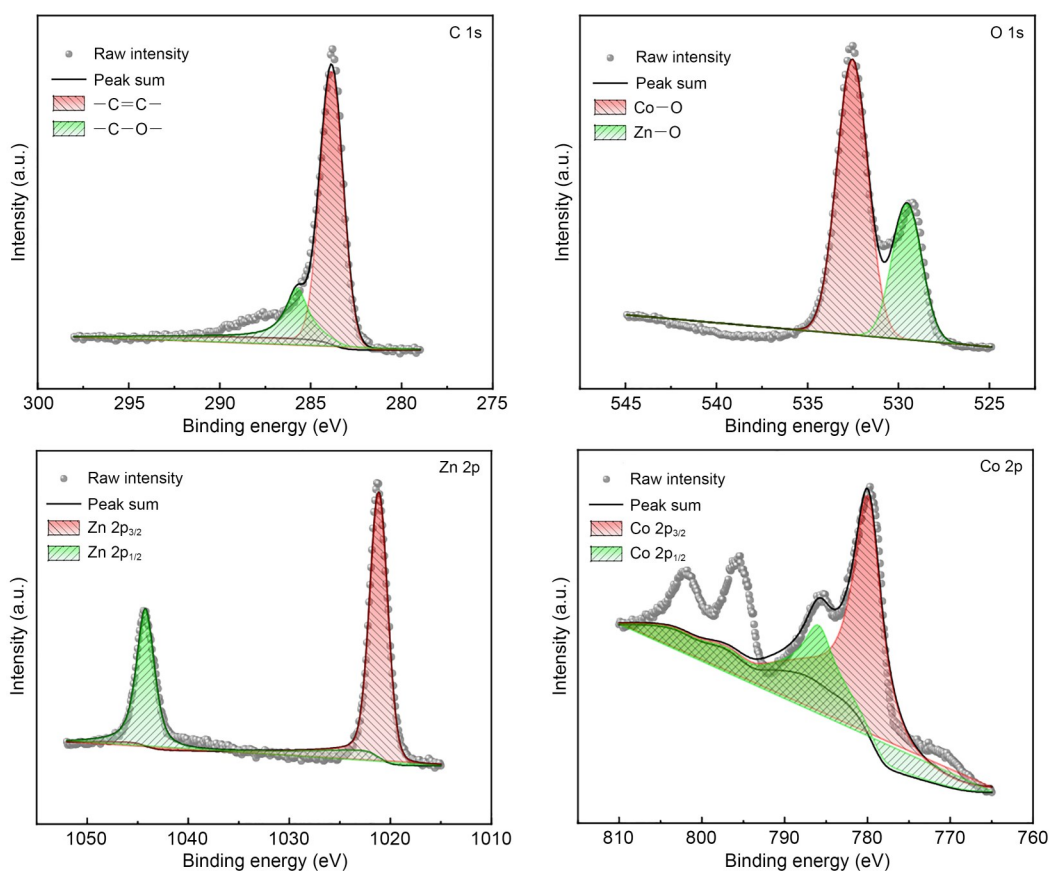


Fig. 5 XPS spectra of $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$

and ZnO lattice oxygen peaks, respectively (Rashid et al., 2014; Zhang QS et al., 2020). Two peaks at 1021.4 ($2p_{3/2}$) and 1044.7 eV ($2p_{1/2}$) were observed for Zn 2p, whereas two peaks at 779.8 ($2p_{3/2}$) and 797.6 eV ($2p_{1/2}$) were observed for Co 2p (Rashid et al., 2014; Zhang QS et al., 2020). Fig. S2 of the ESM shows the full XPS spectra of the three materials; the results indicate that the heterojunction catalyst was composed of four elements, i.e., C, O, Co, and Zn.

3.2 Preparation and performance testing

3.2.1 Doping amount of rGO

Fig. 6 shows the relationship between rGO doping and the RhB degradation rate in $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$. When the doping mass fraction was 0.1%–0.5%, RhB degraded more quickly as the mass percentage of rGO increased, reaching 56.24% in 40 min. We attribute this to the electrical conductivity of rGO, which increases the electron transfer efficiency of $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$, thereby enhancing the catalytic performance. When the rGO doping mass fraction was 0.5%–0.9%,

the rate of RhB degradation declined as the mass fraction of rGO doping increased. This is because too much rGO covered the active sites and reduced the catalytic performance of $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$. The catalytic performance of $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$ was significantly better compared with that of $\text{Co}_3\text{O}_4/\text{rGO}$ and ZnO/rGO under the same rGO doping mass fraction. In addition, the effects of rGO doping mass fraction on the degradation performance of $\text{Co}_3\text{O}_4/\text{rGO}$ and ZnO/rGO were consistent.

3.2.2 Hydrothermal reaction time

The impact of various hydrothermal reaction periods on catalytic activity is shown in Fig. 7. The catalytic properties of $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$ changed with different hydrothermal times. The degradation rate of RhB increased as hydrothermal time increased from 2 to 4 h, and the best catalytic performance was achieved at 4 h, with an RhB degradation rate of 58.78%. At 4–6 h of hydrothermal reaction, the degradation rate of RhB slightly decreased, but exhibited little fluctuation, probably due to the continuous high-temperature and

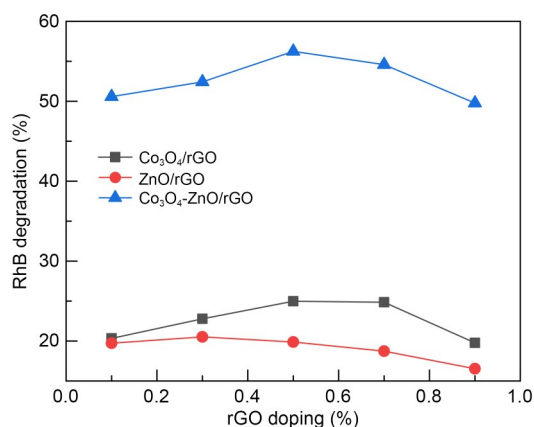


Fig. 6 Effect of rGO doping amount on RhB degradation

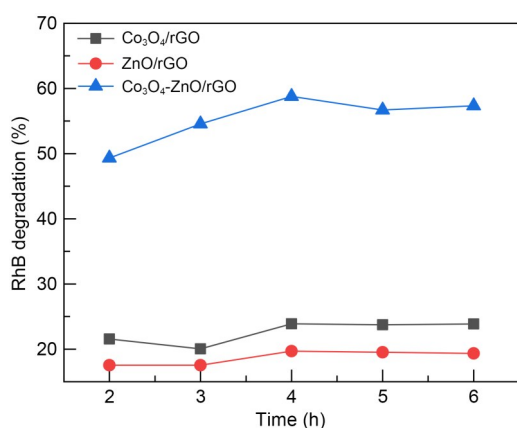


Fig. 7 Effect of hydrothermal reaction time on RhB degradation

high-pressure environment that damaged the crystal structure of Co₃O₄-ZnO/rGO, weakening the catalytic performance.

3.2.3 Calcination temperature

The catalytic performance of Co₃O₄-ZnO/rGO was affected by the temperature range, as shown in Fig. 8. When the calcination temperature was 500–700 °C, the catalytic degradation performance of Co₃O₄-ZnO/rGO was better than that at 300–500 °C, reaching 60.03% in 40 min. The catalytic performances of the products at different temperatures were different because of the different products obtained at the calcination temperatures of the intermediate product Co(OH)₂ that formed during hydrothermal preparation, which can degrade into Co₂O₃ at temperatures below 290 °C, into Co₃O₄ and Co₂O₃ at 290–500 °C, and into Co₃O₄ and CoO at 500–1000 °C. In addition, the small size of the corundum boat used in the calcination process,

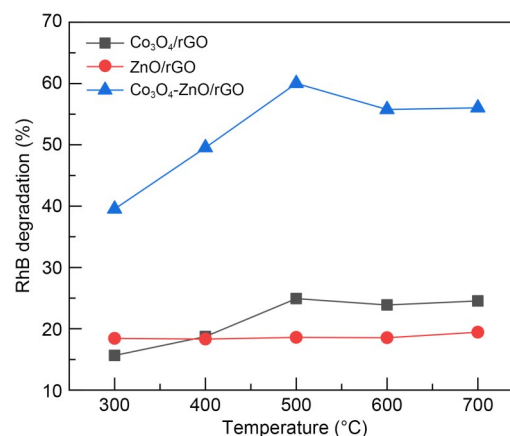


Fig. 8 Effect of calcination temperature on RhB degradation

as well as uneven internal heating, may result in the appearance of multiple products, a phenomenon that affects catalytic degradation performance. When the calcination temperature was 300–500 °C, the increase in the catalytic efficiency of Co₃O₄-ZnO/rGO with temperature was mainly caused by an increase in the yield of Co₃O₄ in the intermediate product Co(OH)₂ by thermal decomposition, and the purity of the prepared Co₃O₄-ZnO/rGO was improved.

3.3 Catalytic performance and stability

We investigated the RhB degradation efficiency of the UV/Co₃O₄-ZnO/rGO/PMS system. First, we tested the impact of adsorption on Co₃O₄-ZnO/rGO performance. About 7.1% of RhB was adsorbed after 40 min. Subsequently, 0.025-g PMS was added to the solution. After shaking for 40 min, the RhB degradation rate increased to approximately 27.55%. This indicates that the addition of PMS improves removal of organic matter from water. Fig. S3 of the ESM shows that the RhB degradation rate of the Co₃O₄-ZnO/rGO/PMS system was better than that of the Co₃O₄/rGO/PMS (25.43%) and ZnO/rGO/PMS (14.56%) systems under the same conditions. The pseudo-first-order reaction rate constants k of RhB degradation were 0.0051, 0.0035, and 0.0022 min⁻¹, respectively. RhB degraded at similar rates in the Co₃O₄-ZnO/rGO/PMS and Co₃O₄/rGO/PMS systems; however, both were faster than the ZnO/rGO/PMS system. This demonstrates that Co plays a key role in the reaction, because its activation effect in promoting the ability of PMS to generate SO₄^{•-} is significantly stronger than that of Zn.

The photocatalytic properties of three catalysts for RhB degradation were tested under visible-light

irradiation. As depicted in Fig. S4 of the ESM, in three systems, $\text{Co}_3\text{O}_4/\text{rGO}/\text{UV}$, $\text{ZnO}/\text{rGO}/\text{UV}$, and $\text{Co}_3\text{O}_4\text{-ZnO}/\text{rGO}/\text{UV}$, the degradation rates of RhB within 40 min were 37.65%, 21.58%, and 57.05%, respectively. RhB degradation in the $\text{Co}_3\text{O}_4\text{-ZnO}/\text{rGO}/\text{UV}$ system had a pseudo-first-order reaction rate constant k of 0.212 min^{-1} , which was 3.36 times faster than that of the $\text{Co}_3\text{O}_4/\text{rGO}/\text{UV}$ system (0.0063 min^{-1}) and 4.32 times faster than that of the $\text{ZnO}/\text{rGO}/\text{UV}$ system (0.0049 min^{-1}). This suggests that photocatalytic degradation of RhB is facilitated by the generation of a $\text{Co}_3\text{O}_4\text{-ZnO}$ heterojunction (Alshaikh et al., 2021). Moreover, the addition of PMS to the system increased RhB degradation efficiency. As shown in Fig. 9, in the $\text{UV}/\text{Co}_3\text{O}_4\text{-ZnO}/\text{rGO}/\text{PMS}$ system, 90.40% of RhB was degraded within 40 min, indicating that the addition of PMS can produce a synergistic effect with the photocatalytic reaction and effectively promote pollutant degradation (Shao et al., 2017; Jin et al., 2020a).

Furthermore, we tested the reusability of the catalyst in the $\text{UV}/\text{Co}_3\text{O}_4\text{-ZnO}/\text{rGO}/\text{PMS}$ system using a five-cycle test. After the fifth reaction, the degradation rate of RhB did not decrease significantly (86.30%) and was only 4.1% lower than that of the first reaction. Fig. S5 of the ESM shows the XPS spectra of the unused catalyst and the catalyst at the end of the reaction. The position and shape of each diffraction peak showed no obvious changes, demonstrating the structural stability of the $\text{Co}_3\text{O}_4\text{-ZnO}/\text{rGO}$. The position of Co changed slightly, while the position and intensity of other elemental peaks did not change significantly. Combined with the repeated use experiment, $\text{Co}_3\text{O}_4\text{-ZnO}/\text{rGO}$ exhibited high stability. The decrease in degradation efficiency was caused by the loss of metal ions by the catalyst.

3.4 Mechanism

3.4.1 Involved active species

We conducted free-radical scavenging studies to identify the free radical species present during RhB degradation. The change in the RhB degradation rate over time following the addition of the inhibitor is depicted in Fig. 10. TBA and MeOH correspond to $\cdot\text{OH}$ and $\text{SO}_4^{\cdot-}$, respectively, and BQ and TEOA correspond to $\text{O}_2^{\cdot-}$ and h^+ , respectively. Injection of MeOH and TBA significantly slowed the process of RhB degradation in the reaction system. The effect of TBA on

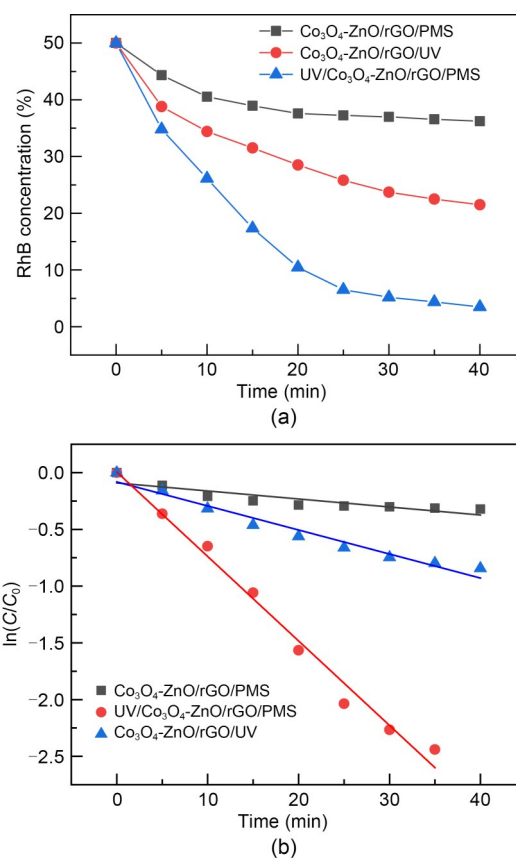


Fig. 9 Degradation of RhB in different systems: (a) RhB treatment efficiency; (b) pseudo-first-order kinetic plot. C is the RhB concentration at time t , and C_0 is the initial concentration of RhB

the RhB degradation rate was the greatest, causing it to decrease from 91.28% to 57.05%, whereas the rate constant decreased from 0.0773 to 0.0249 min^{-1} , indicating that $\cdot\text{OH}$ was the main reactive group in the reaction process. The addition of MeOH reduced the degradation rate of RhB by 24.55%, indicating that $\text{SO}_4^{\cdot-}$ was the main oxidizing species during the reaction. However, RhB degradation was not significantly affected by the addition of BQ and TEOA to the reaction system, demonstrating that few $\text{O}_2^{\cdot-}$ or h^+ were generated in the system to participate in the reaction.

3.4.2 Band gap width and photoresponse range

UV-Vis diffuse reflectance spectroscopy was used to evaluate the light-trapping performance of $\text{Co}_3\text{O}_4\text{-ZnO}/\text{rGO}$, and the results are shown in Fig. 11a. The band gap edges of $\text{Co}_3\text{O}_4\text{-ZnO}/\text{rGO}$ were all approximately 732 nm. We observed two distinctive characteristic peaks at 590 and 610 nm, which were consistent

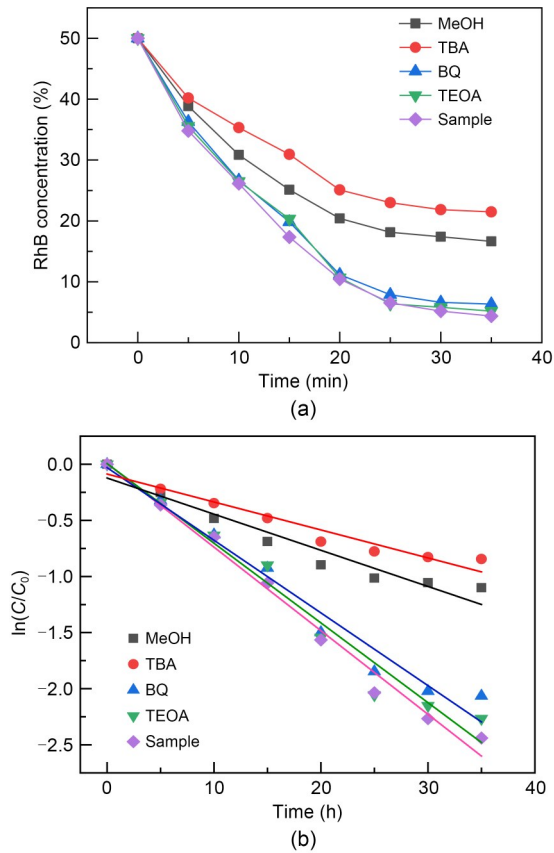


Fig. 10 Effects of active species on RhB degradation: (a) RhB treatment efficiency; (b) pseudo-first-order kinetic plot

with the jump in the tetrahedral coordination structure of the Co (II) ion. Comparing our findings with those of Su et al. (2022), the peak position was slightly red-shifted, which may be because the ZnO in Co₃O₄-ZnO/rGO entered the lattice of Co₃O₄. The relationship between the $(ah\nu)^2$ and $h\nu$ curves is illustrated in Fig. 11b, where a , h , and ν represent the absorption coefficient, Planck's constant, and incident light frequency, respectively. According to Eq. (1), where χ is the electronegativity of the semiconductor (about 5.0 eV) and E_c is the free electron energy on the hydrogen scale (4.3 eV), the Co₃O₄-ZnO/rGO valence band (E_{VB}) can be estimated to be 1.58 eV (Wang et al., 2019). In addition, according to Eq. (2), the conduction band (E_{CB}) of Co₃O₄-ZnO/rGO is -0.18 eV, where E_g is the band gap of the semiconductor energy (1.76 eV) (Zou et al., 2018).

$$E_{VB} = \chi - E_c + \frac{E_g}{2}, \quad (1)$$

$$E_{CB} = E_{VB} - E_g. \quad (2)$$

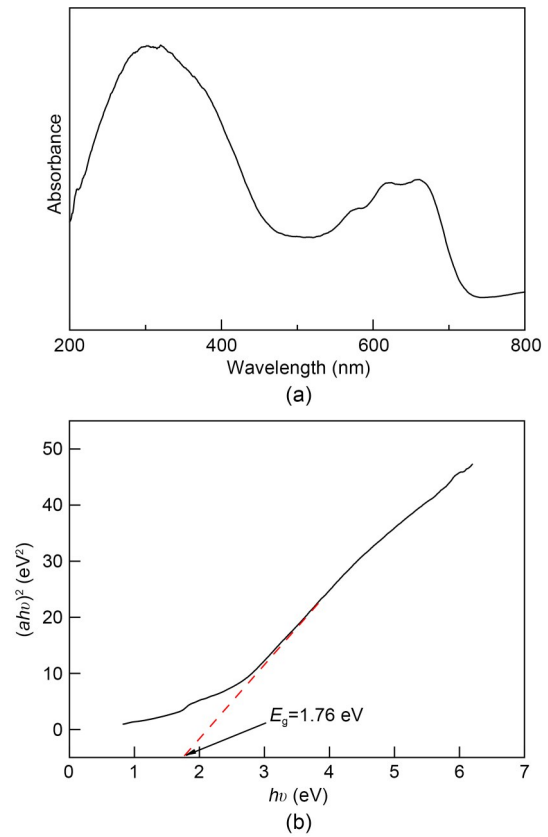


Fig. 11 UV-Vis DRS spectra (a) and $(ah\nu)^2$ and $h\nu$ plots (b) of the samples

3.4.3 Bonding process and Co valence change

Anipsitakis et al. (2005) proved that Co₃O₄ activated PMS to produce SO₄^{•-}. Fig. 12 shows the FT-IR spectrum of the interaction between Co₃O₄-ZnO/rGO and PMS. The stretching vibration peaks of SO₄²⁻ and HSO₃⁻ were located at 1103 and 1249 cm⁻¹, respectively. The Co₃O₄-ZnO/rGO surface hydroxyl peak was located at 3243 cm⁻¹, the Zn-O peak at 483 cm⁻¹, and the Co-O peak at 572 cm⁻¹.

When Co₃O₄-ZnO/rGO was added to the PMS aqueous solution, the intensity of the peak at 1103 cm⁻¹ was substantially higher compared to that of the peak at 1149 cm⁻¹. This indicates that PMS began to decompose, producing SO₄²⁻. The peaks at 1249 and 1103 cm⁻¹ showed blue shifts of 8 and 4 cm, respectively, indicating that the S-O bond was strengthened after Co₃O₄-ZnO/rGO was bonded to the surface of PMS. It is possible that the -OH in S-O-(OH) weakens the attraction to the electron cloud generated by the S-O bond, resulting in the Co (II) of Co₃O₄-ZnO/rGO donating electrons to the -OH groups in PMS.

After we added PMS to the $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$ aqueous solution, the surface hydroxyl peak (3243 cm^{-1}) showed a red shift of 20 cm^{-1} . This indicates that HSO_5^- replaced part of the surface hydroxyl peaks to bond with the surface metal elements of $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$ during the reaction. Simultaneously, the electron-withdrawing effect of Co (III) may have led to its bonding to the hydroxyl group.

Fig. S6 of the ESM shows the XPS patterns of the changes in Co (II) and Co (III) contents in the system before and after the reaction. The molar ratios of Co (II) and Co (III) before the reaction were 35.13% and 64.87%, respectively, whereas those after the reaction were 13.86% and 86.14%, indicating the conversion of Co (II) to Co (III) during the entire reaction process.

The reaction process involves the transfer of electrons through the conversion of different valence states of Co, and the photogenerated carriers can increase the rate of change in the Co valence states. $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$ itself acts as a catalyst to accelerate the reaction process, but its properties do not change. This shows that in the synergistic oxidation reaction system, Co participates in production of $\text{SO}_4^{\cdot-}$, which leads to the conversion of Co (II) to Co (III). Therefore, it can be inferred that photocatalysis increases the yield of $\text{SO}_4^{\cdot-}$ in the $\text{Co}_3\text{O}_4\text{-ZnO/rGO/PMS}$ reaction system, and that the Co (II) or Co (III) created during the $\text{SO}_4^{\cdot-}$ production process further reduces the recombination efficiency of electron-hole pairs.

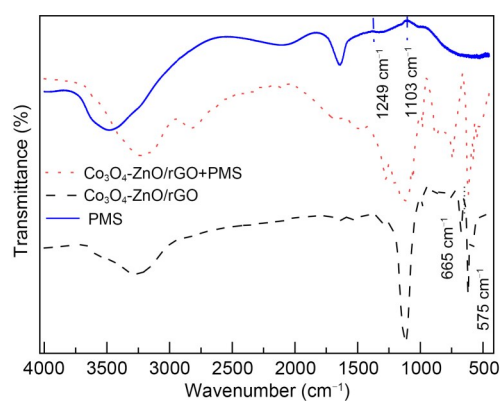


Fig. 12 FT-IR spectra of the interaction between $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$ and PMS

3.4.4 Potential mechanism

Based on the results discussed above, we proposed a mechanism for RhB degradation in the UV/

$\text{Co}_3\text{O}_4\text{-ZnO/rGO/PMS}$ system. First, rGO has a large specific surface area, which offers additional active sites for subsequent reactions, and its electrical conductivity enhances electron-transfer efficiency in the reaction system (Xu et al., 2021). Second, the increased light-harvesting capacity permits $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$ to acquire more photon energy, and heterojunction formation between Co_3O_4 and ZnO reduces the rate at which electron-hole pairs recombine and improves the catalytic ability of the catalyst, which can be also proved from the photoluminescence fluorescence spectrum. Compared with $\text{Co}_3\text{O}_4/\text{rGO}$ and ZnO/rGO , the fluorescence intensity of $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$ is significantly reduced, indicating that the recombination rate of photo-generated electron-hole pairs is reduced and the separation efficiency of photogenerated electrons and holes is improved (Rakibuddin and Ananthakrishnan, 2016; Zhang QS et al., 2020).

As shown in Fig. 13, when $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$ was exposed to visible light, the electrons in the valence band (VB) of both Co_3O_4 and ZnO could be stimulated and transported to their respective conduction band (CB). In addition, under the interference of the built-in electric field, the electrons in the CB of Co_3O_4 were gradually transported to the CB of ZnO, triggering PMS to produce $\text{SO}_4^{\cdot-}$. Similarly, the holes produced by the VB in ZnO were transported to the VB in Co_3O_4 , where they interacted with PMS to generate $\text{SO}_5^{\cdot-}$, which in turn generated $\text{SO}_4^{\cdot-}$ (Jin et al., 2020b; Velmurugan et al., 2021). $\text{SO}_4^{\cdot-}$ can also react with water to form $\cdot\text{OH}$. This effectively reduces the binding of electron-hole pairs and induces more electrons to participate in the degradation process, thereby increasing the reaction rate. Moreover, Co_3O_4 exhibited excellent PMS activation performance, generating more $\text{SO}_4^{\cdot-}$ in the solution under the action of Co (II). Importantly, the free electrons in the solution also accelerated the cycle of Co (II) and Co (III). Finally, the UV/ $\text{Co}_3\text{O}_4\text{-ZnO/rGO/PMS}$ system degraded RhB into CO_2 and H_2O under the action of the main active substances $\text{SO}_4^{\cdot-}$ and $\cdot\text{OH}$.

5 Conclusions

(1) $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$ heterojunction catalysts were synthesized using a hydrothermal method. The distinct particle/sheet structure of $\text{Co}_3\text{O}_4\text{-ZnO/rGO}$ offers a large surface area with numerous active sites for the

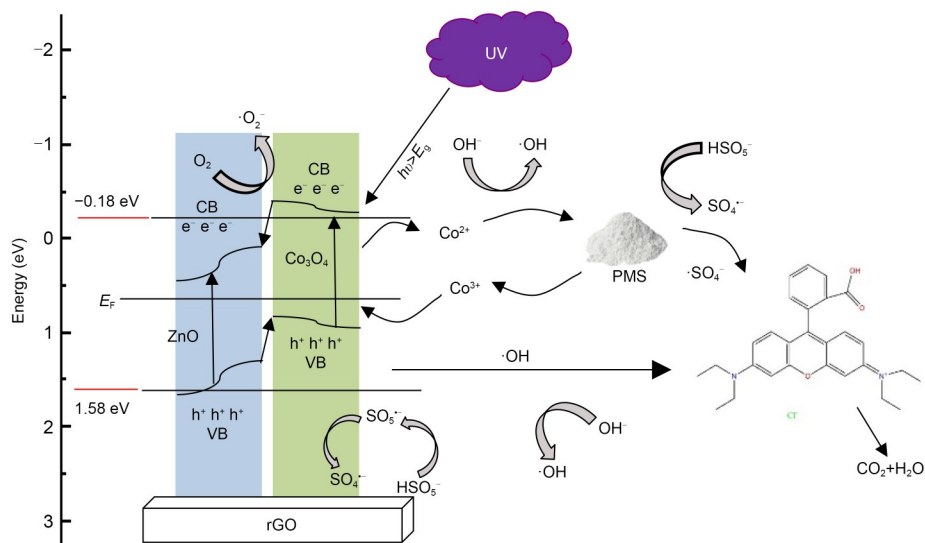


Fig. 13 Potential mechanism of RhB degradation in the UV/Co₃O₄-ZnO/rGO/PMS system (E_F : Fermi level)

reaction, which enables the extension of the carrier lifetime and speeds up the transport and separation of electron-hole pairs, resulting in superior photocatalytic efficiency and the ability to activate PMS compared to stand-alone Co₃O₄ or ZnO.

(2) PMS and photocatalysis exhibit good synergy. PMS better suppresses electron-hole pair recombination by capturing electrons, and the photogenerated carriers generated by illumination also strengthen the redox cycle between Co (II) and Co (III), resulting in more active radicals.

(3) Co₃O₄-ZnO/rGO exhibited a good ability to trigger PMS for RhB degradation in the SR-photo system. In the presence of major radicals, such as SO₄^{•-} and •OH, the UV/Co₃O₄-ZnO/rGO/PMS system cleared 90.4% of the RhB in 40 min.

(4) This study may offer a viable solution for water pollution remediation as well as a reference for designing more effective heterojunction catalysts based on a combination of PMS and photocatalytic technology. Further work is needed to investigate the performance of Co₃O₄-ZnO/rGO materials for visible-light-driven catalytic degradation of organic pollutants in natural water bodies.

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Author contributions

Zhanmei ZHANG designed the research and revised and edited the final version. Xilin CHEN and Ziran HUANG

processed the corresponding data. Yi ZHANG wrote the first draft of the manuscript. Zuqin ZOU and Huaili ZHENG helped to organize the manuscript.

Conflict of interest

Zhanmei ZHANG, Yi ZHANG, Xilin CHEN, Ziran HUANG, Zuqin ZOU, and Huaili ZHENG declare that they have no competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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Electronic supplementary materials

Figs. S1–S6